

## Three decades of remote sensing habitat mapping, along the Egyptian Coast of Aqaba Gulf, Red Sea

Kareem F. Darweesh<sup>1\*</sup>, Ahmed M. Hellal<sup>1</sup>, Samy A. Saber<sup>1</sup>, Sameh B. EL-Kafrawy<sup>2</sup>  
Ali A. H. Abdelsalam<sup>2</sup> and Hussein A. EL-Naggar<sup>1</sup>

<sup>1</sup> Zoology Department, Faculty of Science, Al-Azhar University, Cairo, Egypt.

<sup>2</sup> Marine Sciences Department, National Authority for Remote Sensing and Space Sciences, Egypt.

\*Corresponding Author: kareemfarouq@azhar.edu.eg

### ARTICLE INFO

#### Article History:

Received: Dec. 2, 2021

Accepted: Dec. 17, 2021

Online: Dec. 29, 2021

#### Keywords:

Change detection;

Remote sensing;

Habitat mapping;

Tourism;

Overfishing;

Dahab;

Nuweiba'a;

Red Sea;

Egypt

### ABSTRACT

Over the last three decades, the Egyptian Red Sea shorelines had seen significant growth. In this short period, the Gulf of Aqaba's coastline has been turned into a lengthy belt of tourism settlements and hotels. Due to scarcity of information about this area, herein, a first-time ultimate remote sensing time-series change detection in the region. Landsat satellite data collected over the Gulf of Aqaba during the period from 1985 to 2020 and analysed to estimate urban expansion and shoreline changes in the study area, in addition, from 1995 to 2020 to monitor changes in the coral reef and benthic habitats coverage over time. Landsat 5-TM, Landsat 7-ETM+, and Landsat 8-OLI imagery were used to create a time series of satellite observations. The results showed that the urban development had a large scale increase in the past three decades by about 4.1 km<sup>2</sup> and 2.8 km<sup>2</sup> at Dahab and Nuweiba'a sectors respectively. While shoreline changed on a smaller scale than the urban development in a regular way without harm to natural resources. Changing – urban and shoreline – came along with tourism intensity. Concerning key environmental factors results of Time Series Analyses showed high variance in time especially during the period from 1993 to 2003, while the sea surface temperature had a stable trend. Benthic habitat mapping at Dahab showed stability of the coverage area of the most benthic habitats except for an increase of both corals and macroalgae between 1995 and 2005, whereas there is a fluctuation in coral coverage at Nuweiba'a area over time.

### INTRODUCTION

Sinai Peninsula is in Egypt's northeast, surrounded by the Red Sea, Gulf of Aqaba from the east, Gulf of Suez from the west direction, and by the Mediterranean Sea from the north direction (Sherif *et al.*, 2016). Sinai is known by its diverse ecosystem, biodiversity, dry climate, and distinctive natural environment (Baldwin *et al.*, 1988). Abou Gallum, Saint Catherine, Nabq, Taba, and Ras Muhammed are five natural protectorates in south Sinai with many indigenous coral reefs and species, mainly in the Gulf of Aqaba. Furthermore, the Egyptian Environmental Affairs Agency (EEAA) which is linked with the Ministry of Environment protects the whole coast of Aqaba. The gulf is about 300 kilometers long, stretching from the Protectorate of Ras Mohammed in the south to Taba in the north. The tourist economy in south Sinai, received a lot of attention, so, many hotels have been built along the gulf, resulting in Bedouins didn't have an opportunity to fishing in the front of the hotels back then (Basha, 2017). In the 1990s and 2000s, the tourism business thrived and the natural resources were over exploited and became endangered (Ali, 1998).

On the other hand, Gulf of Aqaba is experiencing an increase in anthropogenic disturbances, mostly due to the increases in the tourism (Rinkevich, 2005 and Naumann *et al.*, 2015). Intensive diving and snorkeling, coastal development, and eutrophication have all been reported to have negative effects on the health of northern Red Sea coral reefs in past studies (Zakai & Chadwick-Furman, 2002; and Naumann *et al.*, 2015). Increases in the cover of turf algae and macroalgae have been recorded in short-term studies in the region (1–3 years), which were positively connected with ambient organic manures (Bahartan *et al.*, 2010 and Naumann *et al.*, 2015).

Dahab is a city on the South Sinai Peninsula; it is a part of the Nabq Protectorate and was once a Bedouin community whose livelihood was based on fishing and herding. The city of Dahab's urban fabric, development pattern, city planning, and demographical structures have all been changed by a 1997 plan aimed to change the source of revenue to a tourism-oriented place and the principal activities taking place in the city. Dahab's economy has shifted from fishing to relying entirely on the tourism sector (MOUNCHU, 1993; Ali, 1998 and Sherif *et al.*, 2016). On the other hand, Nuweiba'a is one of Egypt's most distant outlying with isolated locations and located on the Gulf of Aqaba. It covers 6,977 square meters, or 22.31 percent of the entire area of south Sinai. Nuweiba'a is a tiny coastal community in south Sinai recognized for its real Bedouin culture and pure scenery. After the liberation of Sinai, plans were made to begin developing the entire peninsula; however Nuweiba'a was not completely planned and did not receive much attention (Basha, 2017).

Land cover change has been detected and monitored at various scales using satellite remote-sensing methods by some authors such as Wilson *et al.* (2003) and Hedley *et al.* (2018). Remote sensing observations throughout time are useful for observing and evaluating changes in urban landscapes and land use as development (Batty & Howes, 2001; Herold *et al.*, 2003; Alberti *et al.*, 2004; Xiao *et al.*, 2006 and Ampoua *et al.*, 2018). Remote sensing when paired with GIS and studies involving global positioning systems (GPS), is a strong technique for assessing change in land and coastal substrates (Muller & Zeller, 2002).

To identify regions of disturbance and assess the areal extent and spatial pattern of change, change detection algorithms are applied (MacLeod & Congalton, 1998). To illustrate change, some change detection algorithms adopt simple image overlays, while others use image difference. For measuring change over time, a prominent quantitative technique uses sequences of thematically identified photographs (Singh, 1989). To date, nothing is known about the amount of coral cover in the Red Sea. To characterize distribution, previous research had used qualitative density metrics such as 'low, medium, and high' coral density (Zainal *et al.*, 1993 and Ahmad & Neil, 1994).

Habitat mapping is a spatial representation of how to describe and identify the various habitat components (Valentine *et al.*, 2005). The features of spatial mapping and the distribution of habitats are essential for the analysis and management of the coastal and marine environment (Hamel & Andrefouet, 2010 and Coggan & Diesing, 2011). Various geospatial technologies might be used to generate maps of benthic habitat (White & Fitzpatrick, 2007 and Australia, 2012). The capability to characterize the seabed and its related ecosystems has improved as GIS and remote sensing tools have improved (Williams *et al.*, 2010 and Brown *et al.*, 2011). Use of remote sensing can help with both appropriate sample site selection and mapping the distribution of benthic life (Blakey *et al.*, 2015 and Zakaria *et al.*, 2019). Previously, satellite imaging had been applied to detect benthic cover in shallow water in the marine environment (Lucas & Goodman, 2014). Remote sensing plays an important role in such approaches, hence a thorough understanding of the spectral features of organisms (Swift & Bower, 2003).

Key environmental factors include light regime is among the most important environmental aspects, wave activity, long water residence durations in lagoons with restricted tidal exchange, as well as wave and tidal-driven current flows on fore-reefs, can all have a major impact on bleaching occurrence patterns and changes in water turbidity, as well as other environmental variations can affect coral resilience (Coles & Jokiel, 1978; Nakamura & van Woesik, 2001; Nakamura *et al.*, 2003; McClanahan *et al.*, 2005 and Anthony *et al.*, 2007)

In this study, The Egyptian waters of the Gulf of Aqaba have been subjected to quantitative evaluations of coral reef habitat over the past three decades during the period from 1995 to 2020 and change detection in urban development and shoreline takes place from 1985 to 2020. The major purpose of this study is to use remote sensing techniques to (1) analyse anthropogenic effects on shoreline and urban development changes in the study areas and (2) identify benthic habitat changes over the previous decades in conjunction with effective environmental factors.

## MATERIALS AND METHODS

### 1- Study Area

Two sectors were selected along the Egyptian coast of Gulf of Aqaba to express the main human activities (tourism and fishing). The first sector is Dahab area which represents the tourism activity and extended about 43.3 Km, and the second sector is Nuweiba'a area which represents the fishing activity and extended about 48.5 Km (Fig. 1).

### 2- Field survey

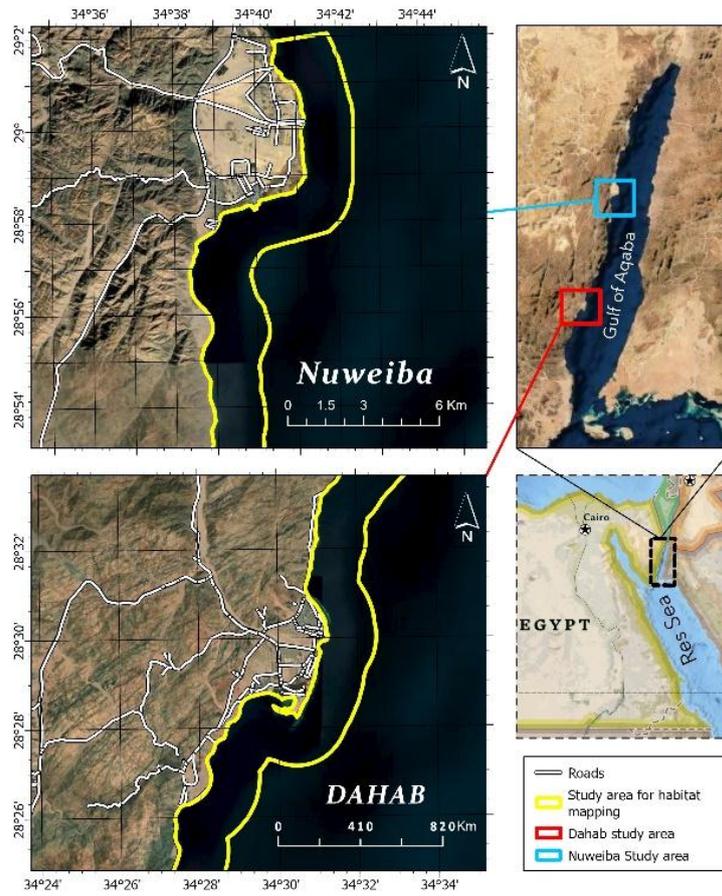
Field survey was planned to represent the range of physical environments within the study area. The locations of approximately 90 points (Fig. 2) were recorded for all water body habitat types as a guide for multispectral classification. Ground truth points at the study area were conducted over a period of 6 days during August 2020 using handheld Trimble GPS inside waterproof case. At each site the handheld GPS unit to pilot the boat to the pre-programmed waypoint of the site's coordinates was used (English *et al.*, 1997).

### 3- Remote Sensing Data

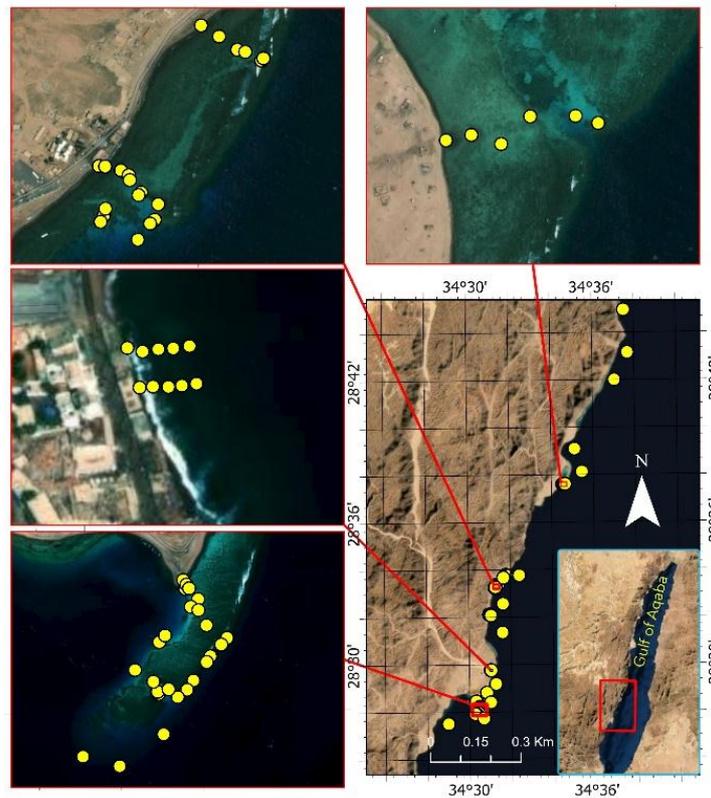
Five satellite images were collected by the Landsat series of sensors (Landsat 5-TM, Landsat 7-ETM, and Landsat 8-OLI) were used to evaluate urban development, shoreline, and marine habitat changes during the years 1985, 1995, 2005, 2015, and 2020. Images were downloaded from the United States Geological Survey (USGS) (<http://earthexplorer.usgs.gov/>). Each of the Landsat sensors had slightly different spectral band combinations (Table, 1). All satellite images were georeferenced and geometrically corrected to match a WGS 84 datum (world geographic system), UTM (Universe Transverse Mercator) Projection with Zone 36 North. Images were radiometrically and atmospherically corrected, and a water-column correction was applied. All images were processed using the ENVI 5.3 software.

**Table 1.** Landsat sensors and spectral bands.

Band	Name	Landsat 8 OLI TIRS + $\mu\text{m}$ )	Landsat 5 and 7 TM and ETM + ( $\mu\text{m}$ )
1	Ultra blue (coastal band)	0.433–0.453	-----
2	Blue band	0.450–0.515	0.45-0.52
3	Green band	0.525–0.600	0.52-0.60
4	Red band	0.630–0.680	0.63-0.69
5	Near Infrared (NIR)	0.845–0.885	0.76-0.90



**Fig. (1).** Landsat satellite image showing the locations of the study sectors.



**Fig. (2).** Landsat satellite image showing locations of ground truth data.

#### 4- Habitat Mapping

##### **Image pre-processing:**

The data collected from Landsat satellites sensors included significant amount of information that is not reflectance from the area of interest because of the complex pathways in the atmosphere and water column. Only about 10 % of the data received comes from the ocean bottom (**Mishra *et al.*, 2005**). In fact, atmospheric scattering accounts for the majority of the signal received by the sensor (**Robinson, 2004**). After compositing satellites bands, the resolution enhancement took place by pan-sharpening, the purpose of pan-sharpening was to spectrally sharpen low spatial resolution image data with high spatial resolution image data figure. The 5-band multispectral ETM and TM low-resolution (30 m) have been enhanced to (15 m) with ENVI modules. Also resampling achieved for TM sensor to reach the unified resolution.

Atmospheric correction computed for all images using dark object subtraction (DOS). ENVI software was used for performing the DOS correction as it is an automated process which produces a corrected multispectral image. Sun glint removal technique was applied to the atmospherically corrected and masked subset. A sample of image pixels was selected from multiple sites of optically deep water featuring uniform sun glint. This process was performed according to **Hedley *et al.* (2005)**. Water column correction technique was performed on the atmospherically corrected satellite images subset after the sun glint correction, sun glint and water column correction implemented using ENVI 5.3 software.

##### **Image classification:**

Satellite Image Supervised classification techniques require the analyst to specify the types of ground cover in a scene through the use of training data (**Lillesand *et al.*, 2004**). The generation of a classification has two distinct steps (training and classification). Training is the process of setting a spectral envelope for a class and, for supervised classification, requires a priori information about the image data and habitats to be mapped (**Green *et al.*, 2000**). Maximum Likelihood Classifier is one of the most popular methods of categorization in remote sensing technique, where the pixel with the maximum likelihood is classified into the corresponding class (**Richards and Ervin, 2008**).

In this study, the image classified to 5 classes: (coral, sand, macro algae, seagrass and deep water). Supervised classification done by using 90 field ground truth points representing all these classes. Training points converted to defined region of interest (ROI) on satellite image in the same date of field survey. ROI data converted to spectral library depending on one date validated to be used in other dates and depending on spectral signature to each class. The supervised classification performed by maximum likelihood algorithm using ENVI 5.3 software.

#### 5- Change detection analysis

The benthic habitat change detection analysis was performed using the results of the supervised classification of four satellite images from (1995-2020). Results of the supervised classification were converted into vector layers. (QGIS) software used to perform a contextual editing for each class to complete the calculation of class's areas. The difference

in the area of each class in sequential satellite images was then computed. From 1985 -2020 five satellite images were used for classification are used for urban and shoreline digitizing.

## 6- Key environmental parameters

The data of key environmental parameters (sea level rise, turbidity and sea surface temperature) were obtained for study area during the period from 1993 to 2020. Diffuse water column attenuation coefficient ( $KD\ 490\ m^{-1}$ ) was obtained from the Goddard Earth Sciences Data Information Services Centre's (GES DISC). High-resolution blended analysis of daily sea surface temperature (SST) ( $^{\circ}C$ ) was downloaded from the NOAA Earth System Research Laboratory Physical Sciences Division (<https://www.esrl.noaa.gov/psd/>). Daily sea level was obtained from the National Aeronautics and Space Administration Jet Propulsion Laboratory NASA JPL (<https://sealevel.nasa.gov/data-analysis-tool/>).

## RESULTS AND DISCUSSION

### 1- Urban development and shoreline changes

Results of changes in urban area during the 35 years are presented in **Figures (3-5)**. The data clearly showed that the total changes in the urban area reached + 4.1 km<sup>2</sup> and 2.8 km<sup>2</sup> at Dahab and Nuweiba'a sectors respectively. However, the urban area in the year 1985 was about 0.48 km<sup>2</sup> at Dahab sector then it markedly increased during the following successive years reaching to 4.583 km<sup>2</sup> in the year 2020. While at Nuweiba'a sector, the urban area was 0 km<sup>2</sup> in the year 1985 after that it exhibited the same pattern as in Dahab area reaching to 2.818 km<sup>2</sup> in the 2020.

Changes in the shoreline occurred primarily due to man-made development (accretion) or natural causes (erosion). Total changes in land area along the shoreline from 1995 to 2020 at Dahab sector was about 0.36 km<sup>2</sup> by which, the total rate of changes by accretion was 0.19 km<sup>2</sup> and by erosion rate was 0.17 km<sup>2</sup>. While at Nuweiba'a sector, the changes in the shoreline were mainly due to the erosion rate (0.45 km<sup>2</sup>) if compared with the accretion rate (0.01 km<sup>2</sup>) (**Figs. 6 & 7**). The results of the present study reflect the succession of political events and economical phases change, which had a huge impact on the coastal and urban structure development at Dahab and Nuweiba'a areas during the past four decades.

Between 1980 and 1988, Bedouins began to settle in southern Sinai area (**Lavie, 1991**); in the meantime, the Egyptian Government began to play a role in the development of Sinai (**Ali, 1998**). According to Law 3 of 1982, the process of orienting urban growth based on land use, services, infrastructure, and other factors to provide an efficient living environment was described as urban planning (**MOHPUC, 1997**). To nationalize Sinai, the Government encouraged the tourist industry to promote internal migration (**Hemdan, 1980** and **Karkabi, 2013**). The Bedouins benefited financially from tourism, but it also had an impact on their culture, since it led them to settle and leave their nomadic lifestyle. Some Bedouins migrated to coastal tourist towns, South Sinai became a popular tourist destination, and the built environment, including residential units, expanded. Many career possibilities arose as a result of the tourist industry's development. Europeans, also, sought for chances that provided a higher quality of life. "Tourism migration" and "lifestyle migration" are two terms used to describe this type of migration. In 1980s, as a result in the increases interest in

marine life and the shortage of local specialists in that sector the Europeans in their twenties and thirties moved to Sinai to pursue careers in such fields (**Karkabi, 2013**). The tourism industry grew dramatically between the year 1988 and 2005 (**Karkabi, 2013**). By the 1990s, the number of foreign migrants had grown to the point that the Government had to adopt residential legislation for them (Law 230 of 1996) (**Ali, 1998**).

After 2011 Egypt's domestic and international political concerns were unsteady, affecting the country's economic status and its significant progress. Political events slowed the growth of Sinai, which was relied on the tourism business (**Sherif et al., 2016**). On the other hand, Nuweiba'a city is seeing rapid development, particularly in the construction. The Egyptian Government is willing to provide critical capacity building in the elements of public housing, services, and infrastructure; however, this is no longer the case. The majority of these projects, on the other hand, have been in action since the 1990s in order to attract resources and manpower to the newly built urban area (**Basha, 2017**).

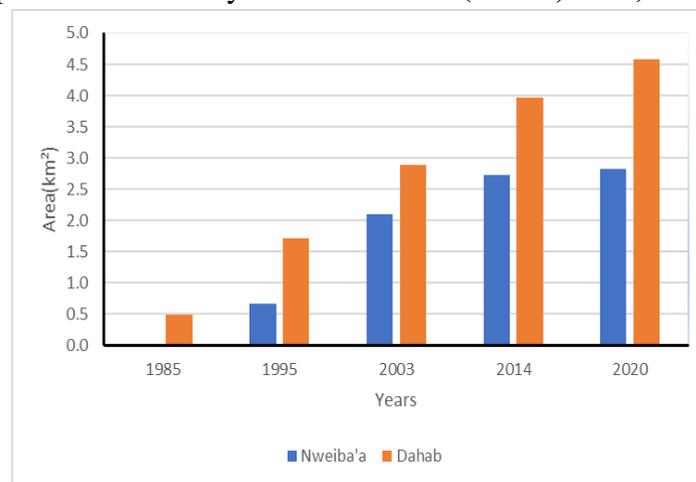


Fig. (3). Changes in urban area (km<sup>2</sup>) at Dahab and Nuweiba'a sectors during 1985-2020.

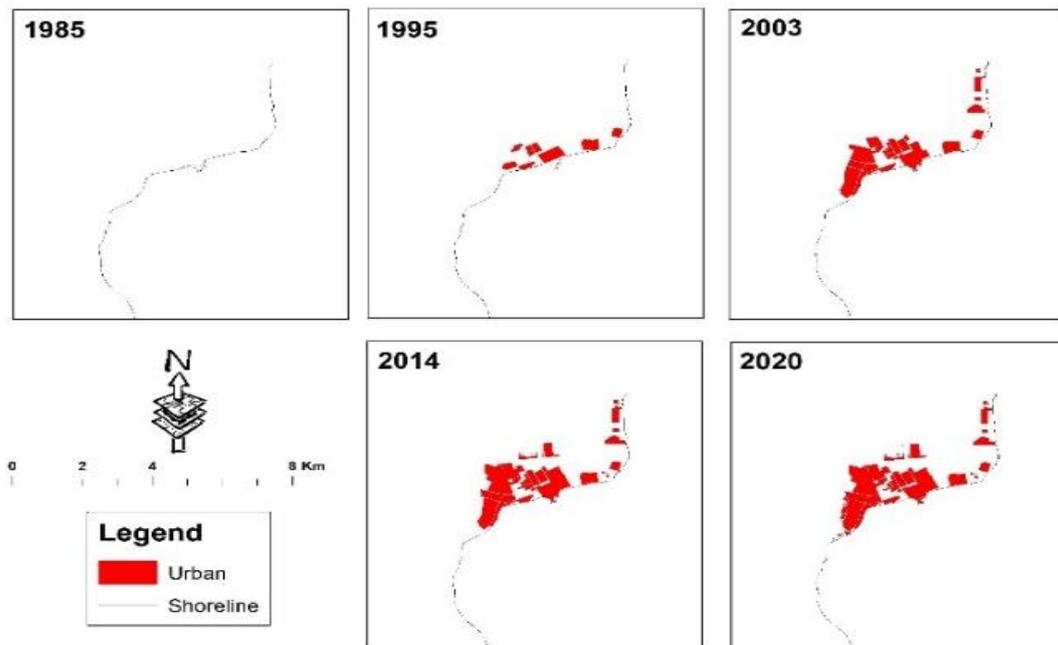


Fig. (4). Changes in urban area at Nuweiba'a area during the period from 1985 to 2020.

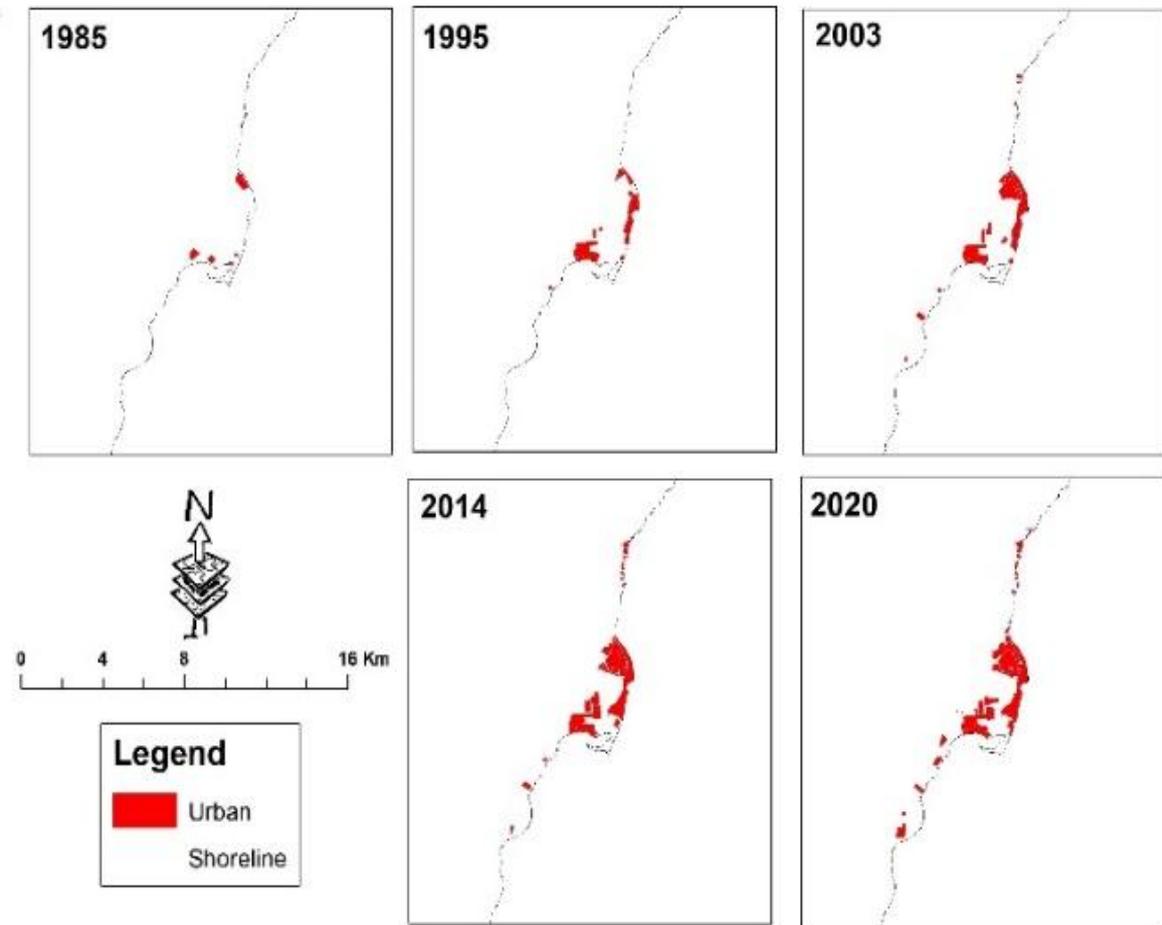


Fig. (5). Changes in urban area at Dahab area during the period from 1985 to 2020.

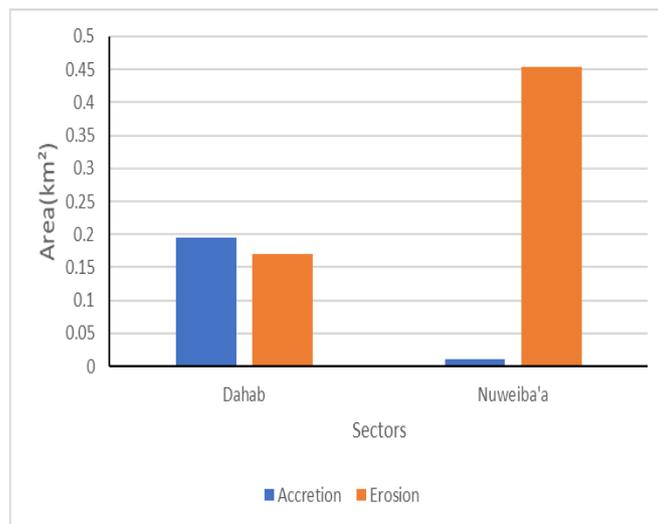
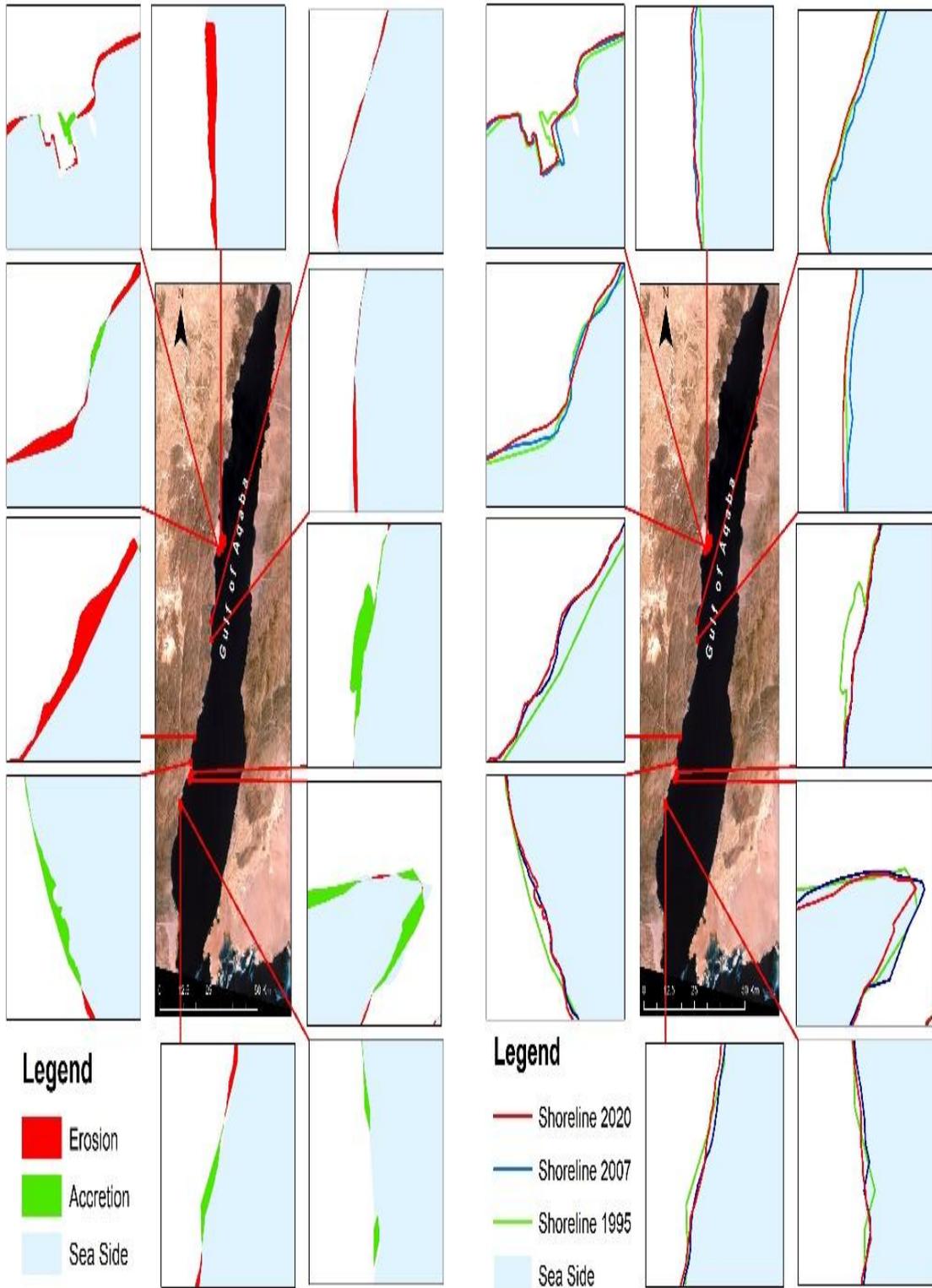


Fig. (6). Changes in shoreline area (km<sup>2</sup>) at Dahab and Nuweiba'a sectors during the period from 1985 to 2020.



**Fig. 7. Changes in the shoreline at Nuweiba'a and Dahab sectors during the period from 1985 to 2020.**

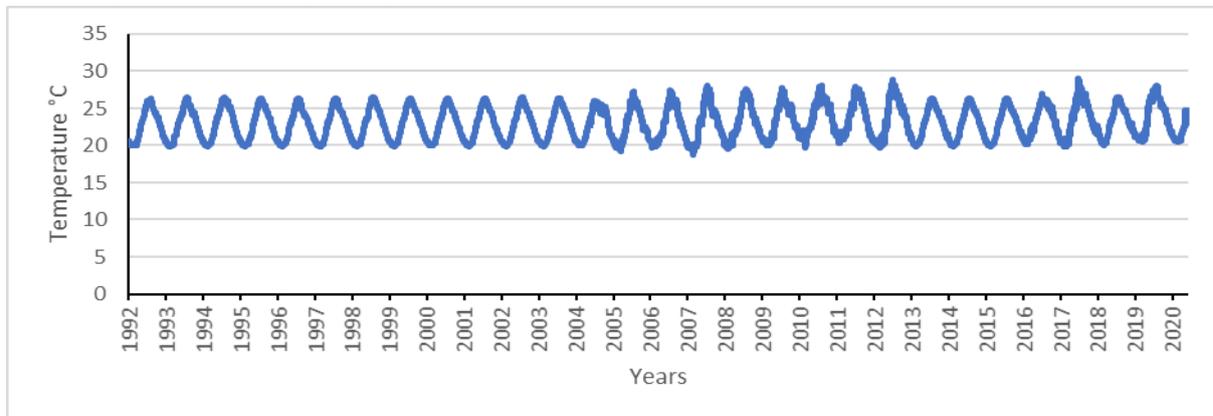
## 2- Time- Series Analysis of key environmental factors changes:

Results of Time Series Analyses of the averages sea surface temperature during 27 years (1993- 2020) showed a stable trend during summer (peak) and winter (bottom) seasons (**Fig. 8**). Sea surface temperature increasing in larger scale such as oceans can be potentially related to bleach events models known as “time to reef extinction had been driven” (**Hoegh-Guldberg, 1999** and **Sheppard, 2003**). Generally, it is well known that coral populations living in warmer seas can tolerate greater temperatures than those living in colder waters of the same species (**Coles *et al.*, 1976**). Moreover, based on **Berkelmans & Willis (1999)** observations **Buddemeier & Fautin (1993)** suggested that corals may be able to acclimate (a physiological response at the individual level) or adapt (a genetic response at the population level) to changing in thermal regimes. According to **Buddemeier & Fautin (1993)** opinion, the shifts in coral host-dinoflagellate relationships, resulting in “different” ecospecies of corals have high tolerances against temperature with tolerant strains of zooxanthellae.

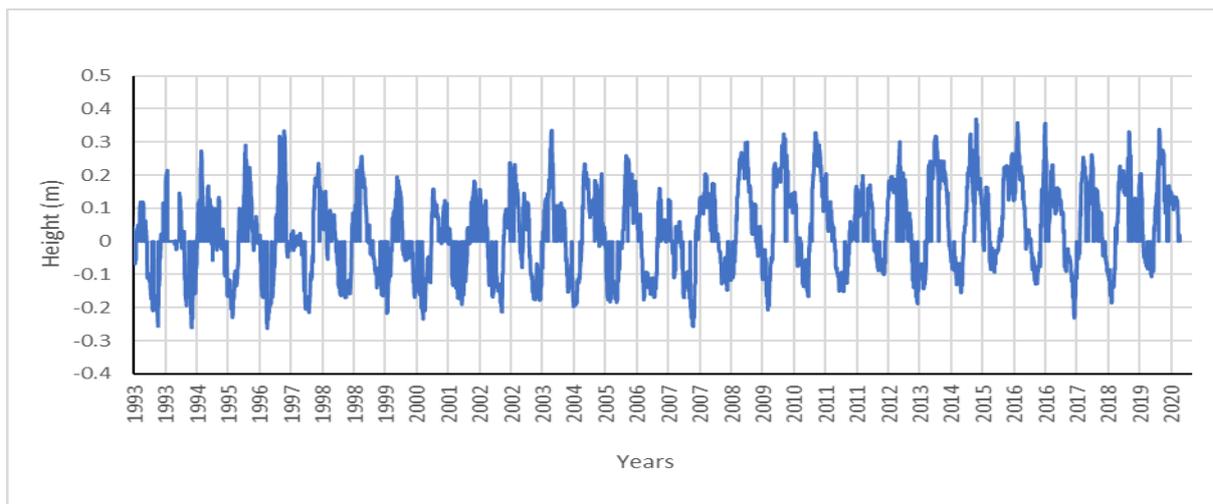
Regarding to the sea level which indirectly refers to the tidal movement, in the present work, the obtained results of the sea level analyses (low tide) showed that there is markedly decline in sea level with long hours exposure during the period from 1993- 2003. After that it took the normal trend of rise and fall curve (**Fig. 9**). The moon and sun's gravitational forces influence not only the astronomical shelf and ocean surface tides, but also the flow and flood tidal currents. When tides interact with the seabed and shelf structures (coral reefs and atolls), friction, vertical fractures, turbulence, eddies, and energy transfer to other processes like internal waves and fronts occur (**Dixon, 2011**). It's difficult to link the highest limit of coral development to a single tide level since so many things influence it. A better description would be a range between mean low water springs and mean low water neap tides (**Scoffin, 1977** and **Hopley, 1986**). In many tropical regions impacted by rainy seasons, mean low water springs vary significantly by season; for example, mean low water springs in the Red Sea and Gulf of Aqaba vary by 0.5 m between winter and summer monsoons (**Dixon, 2011**). The timing of local severe spring low water is crucial for coral reef upper development, i.e., when and how long they are exposed to air and direct UV sunshine, which can lead to coral desiccation. In **1986**, **Hopley** mentioned that corals exposed to air and direct sunlight for more than 3 hours begins to wilt and expel their symbiotic algae zooxanthellae. However, in assessing coral development, exposure durations, and desiccation, the contrast between local diurnal and semi diurnal tidal regimes is essential (**Dixon, 2011**).

The obtained data of turbidity (KD490) showed that there is an elevation rate during the period from 1997 to 2002, and then it gradually decreased till 2020 year except during 2012 which was relatively higher (**Fig. 10**). Turbidity is a key water quality parameter that represents the amount of light absorbed or scattered in the water column by suspended particulate matter (**Van Duin *et al.*, 2001** and **Flores *et al.*, 2012**). Despite occupying 30 % of reefs in the coral triangle and 12 % of reefs globally, turbid coral reefs are relatively unexplored (**Sully & Van Woesik, 2020**). The lack of data on turbid reefs is largely due to logistical issues associated with working in low visibility conditions both directly (in situ) and indirectly using remote sensing technologies (**Morgan *et al.*, 2020**). This has resulted in a poor understanding of how these reefs function, from the individual coral to the reef ecosystem (**Zewlifer *et al.*, 2021**).

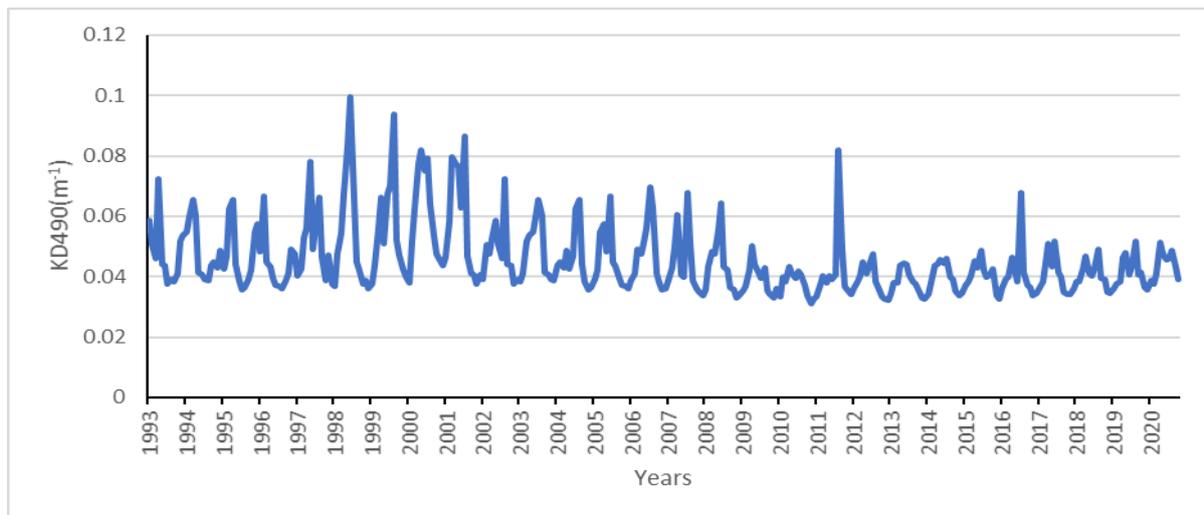
Suspended sediments have traditionally been thought to have a strong influence on coral reef health (Larcombe *et al.*, 2001; Sanders & Baron-Szabo, 2005; Palmer *et al.*, 2010 and Weber *et al.*, 2012) and coral resilience (Rogers, 1990; Gilmour, 1999 and Fabricius, 2005). Over the last 20 years ago, several studies have revealed high coral cover in turbid reefs, implying that turbid coral communities may be able to withstand natural marginal growth conditions, implying that they may be more resilient to both local and global stressors, so that turbid coral reefs as a habitat may become important for coastal conservation (Anthony & Larcombe, 2000; Palmer *et al.*, 2010; Goodkin *et al.*, 2011; Wilson *et al.*, 2011; Browne *et al.*, 2013; Morgan *et al.*, 2016; Pizarro *et al.*, 2017 and Browne *et al.*, 2019). The influence of future global and local environmental disturbances on coral reefs and other habitats till now are still unknown and unexpected. Coral reef resilience to global climate impacts (e.g., warming oceans, cyclones, rising sea levels) will be influenced by the coral communities' ability to adapt these threats. Yet, we have a limited understanding of how these multiple stressors interact with water quality to influence reef function (Zewlifer *et al.*, 2021).



**Fig. (8).** Time-Series Analysis of changes in the averages of sea surface temperature ( $^{\circ}\text{C}$ ) in the Gulf of Aqaba during the period from 1993 to 2020.



**Fig. (9).** Time-Series Analysis of changes in sea level (meter) in the Gulf of Aqaba during the period from 1993 to 2020.



**Fig. (10).** Time-Series Analysis of changes in turbidity – KD490- ((m – 1) in the Gulf of Aqaba during the period from 1993 to 2020.

### 3- Benthic habitat changes:

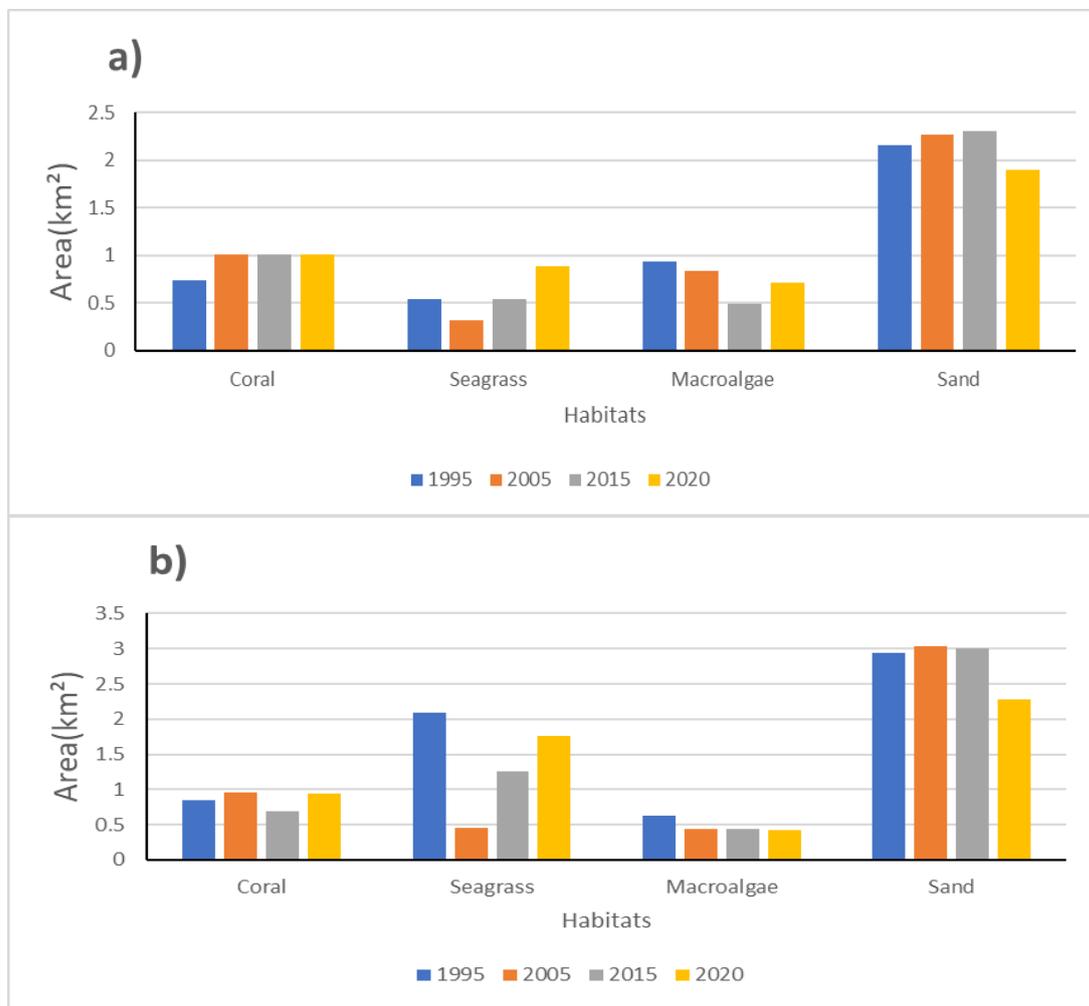
Data obtained from the satellite images at Dahab sector showed that coral reef and seagrass cover in the study area increased by about 0.27 km<sup>2</sup> and 0.35 km<sup>2</sup> respectively during the period from 1995 to 2020. In contrast, macro algae and sand habitat decrease by about 0.23 km<sup>2</sup> and 0.27 km<sup>2</sup> respectively at the same period of study (**Figs. 11a & 12**).

While at Nuweiba'a sector, seagrass, macroalgae and sand habitats decreased by about 0.32 km<sup>2</sup>, 0.2 km<sup>2</sup> and 0.7 km<sup>2</sup> respectively; on the other hand, coral reef habitat slightly increased by about 0.01 km<sup>2</sup> (**Figs. 11 b & 13**). In the present study, results showed that there was an increases area of coral reef during the period from 1995 to 2005, unlike the expected of shrinking as a result from the interference of man with intense activities in this region. Due to scarcity of the information about status of benthic habitat during such period, which lead the authors to suggest the first hypothesis for explanation, and linking the key environmental factors with habitat mapping. The long term neap tide which is happening mainly in summer season at Gulf of Aqaba leaving the whole reef flat including corals and macro algae exposed to intensive solar energy and high UV radiance for long hours of photo period. The authors indicated that due to such conditions may be led eventually to desiccate corals and put macro algae beyond its tolerance range. These findings are agreement with that found by **Hopley (1986)**, **Häder *et al.* (2001)** and **Dixon (2011)**. Also, after that, with the end of such conditions and the rises of sea level corals could flourish faster than macro algae and the competition between them on space is naturally.

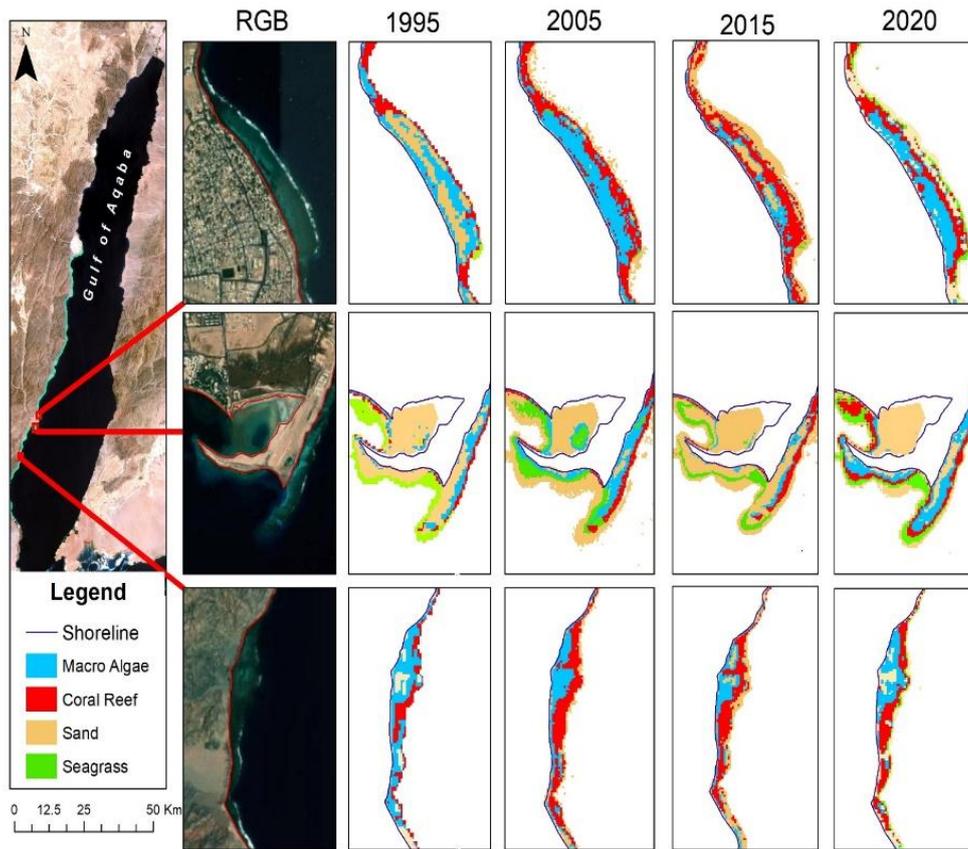
In the present study at Dahab sector, the results also showed that there was stability in benthic habitats especially in coral reef habitat with tourism development during the period of study. Our results confirmed the findings obtained by **Tilot *et al.* (2008)** and **Darweesh *et al.* (2021)** who concluded that there was no relationship between intensity of tourism and coral cover and unlike with the findings obtained by **Khaled *et al.* 2019** and **Abo El Enin *et al.* (2020)** who indicated that the expansion of the touristic urban along Hurgada area was the main reason for coral coverage area degradation.

In their long-term research, **Reverter *et al.* (2020)** reported that hard coral cover did not affected negatively in any of the site groups, although it did significantly grow. However,

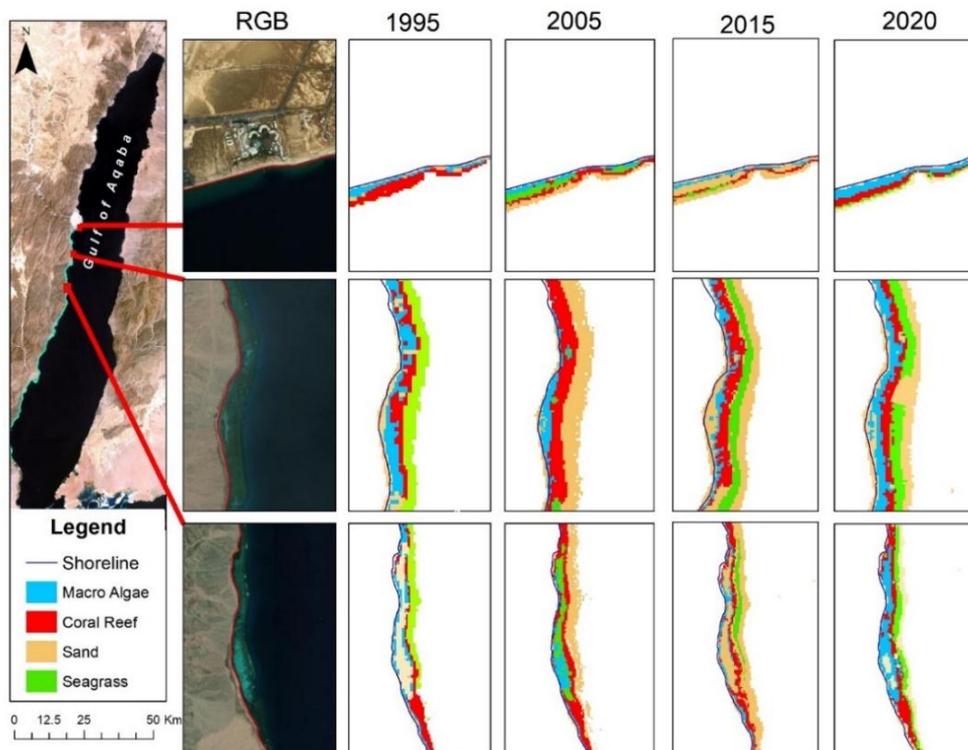
in the present study at the Southern part of the Nuweiba'a region, there was a fluctuation in the hard coral cover over the past decades, although there is low intensity of human impact except for the Bedouin traditional fishing activity. These findings are agreed with that reported by **Sarnowski (2010)** and **Darweesh *et al.*(2021)** who mentioned that the Bedouin fishermen used primitive and destructive fishing methods that could have caused severe damage to coral reef habitats at Nuweiba'a area. Also, in their study **Mona *et al.* (2019)** stated that, the illegal and traditional fishing methods of the Bedouin community are of the most reasons for altering ecosystems at Nabq Protectorate, South Sinai. Because algae are natural space competitors with corals and when herbivores capacity is decreased due to overfishing, algae can spread and affect coral reef functioning, resulting in coral reef degradation and a community phase change leading to increased coral mortality. These findings are in full agreed with many investigators such as **Bellwood *et al.* (2004)**, **Rasher & Hay (2010)**, **Hoey & Bellwood (2011)**, **Dixson *et al.* (2014)**, **Haas *et al.* (2016)** and **Zaneveld *et al.* (2016)**.



**Fig. (11 Changes in benthic habitat areas (km<sup>2</sup>) at Dahab sector (a) and Nuweiba'a sector (b) during the period from 1995 to 2020.**



**Fig. 12. Supervised classification of benthic habitats (coral, seagrass, macro algae and sand) changes between 1995 and 2020 at Dahab sector.**



**Fig. 13. Supervised classification of benthic habitats (coral, seagrass, macro algae and sand) changes between 1995 and 2020 at Nuweiba'a sector.**

## CONCLUSION

Pooled data from the deferent remote sensing changes detection, showed that:

- 1- Urban development had a large scale increasing in the past three decades due to the community change from being dependent on fishing to be a touristic orientation
- 2- Shoreline change in a smaller scale than the urban development in a regular way without harm to natural resources.
- 3- Both above mentioned changing – urban and shoreline- came along with tourism intensity didn't relate to hard coral cover or the benthic habitats.
- 4- A local decreasing in hard coral cover at the hot spots of intensive Bedouin's fishing areas.
- 5- Sea level and turbidity time series analyses showed a high variance in time especially in a period from 1993 to 2003, while the sea surface temperature had a stable trend.
- 6- Habitat mapping at Dahab area showed a stability of the coverage area of most of benthic habitats except for an increasing of both corals and macro algae between 1995 and 2005.
- 7- Nuweiba'a coral reef coverage area revealed a fluctuated coverage may be due to the Bedouin's fishing behaviour.

## ACKNOWLEDGMENT

We thank the National Authority for Remote Sensing and Space Sciences, Egypt. (NARSS) for facilitating the work and express our gratitude especially to Marine Sciences Department Senior, Dr. Mustafa Atef. We also thank Amir Ashraf, Manar Ahmad and Shima'a Refa'at for their help in (GIS). Deep thanks to Emad Abdullah for his help in the field.

## REFERENCES

- Abo Elenin, H.; Saber, S.A.; El-Kafrawy, S.B. and El-Naggar, H.A. (2020).** An Integrated Field Survey and Remote Sensing Approach for Marine Habitat Mapping Along Hurghada Coast, Red Sea, Egypt. *Egyptian Journal of Aquatic Biology & Fisheries*, **24** (7): 287-306.
- Ahmad, W. and Neil, D.T. (1994).** An evaluation of Landsat Thematic Mapper (TM) digital data for discriminating coral reef zonation: Heron Reef (GBR). *International Journal of Remote Sensing*, **15**(13): 2583-2597.
- Alberti, M.; Weeks, R. and Coe, S. (2004).** Urban land-cover change analysis in central Puget Sound. *Photogrammetric Engineering & Remote Sensing*, **70**(9):1043-1052.
- Ali, D.F. (1998).** Case Study of Development of the Peripheral Coastal Area of South Sinai in Relation to its Bedouin Community. (MSc. Thesis of Urban and Regional Planning). Virginia Polytechnic Institute and State University.
- Ampoua, E.E.; Ouillon S.; Iovan, C. and Andréfouët S. (2018).** Change detection of Bunaken Island coral reefs using 15 years of very high-resolution satellite images: A kaleidoscope of habitat trajectories. *Marine Pollution Bulletin*, **131**: 83–95.

- Anthony, K. and Larcombe, P. (2000).** Coral Reefs in Turbid Waters: Sediment-Induced Stresses in Corals and Likely Mechanisms of Adaptation. In Proceedings of the Ninth International Coral Reef Symposium, Bali, Indonesia, **23–27**: 239–244.
- Anthony, K. R. N., Connolly, S. R. and Hoegh-Guldberg, O. (2007).** Bleaching, energetics and coral mortality risk: Effects of temperature, light, and sediment regime. *Limnology and Oceanography*, **49**: 2201–2211.
- Australia, G. (2012).** OzCoasts Australian online coastal information. (<http://www.ozcoasts.org.au>).
- Bahartan, K.; Zibdah, M.; Ahmed, Y.; Israel, A.; Brickner, I. and Abelson, A. (2010).** Macroalgae in the coral reefs of Eilat (Gulf of Aqaba, Red Sea) as a possible indicator of reef degradation. *Mar. Pollut. Bull.*, **60**: 759–764.
- Baldwin, M.; Ferguson, D.; Saterson, K.A. and Wallen, I.E. (1988).** The biological resources of the Arab Republic of Egypt status and recommended conservation needs. USAID Report.
- Basha, M.F. (2017).** Community Resilience for Coastal Peripheral Areas Development in Nuweiba, South Sinai (Doctoral dissertation, University of Stuttgart).
- Batty, M. and Howes, D. (2001).** Predicting temporal patterns in urban development from remote imagery. In: **Donnay, J., Barnsley, M.J. and Longley, P.A. (2001).** Remote Sensing and Urban Analysis. Taylor and Francis, London 169-171.
- Bellwood, D.R.; Hughes, T.P.; Folke, C. and Nystrom, M. (2004).** Confronting the coral reef crisis. *Nature*, **429**: 827–833.
- Berkelmans, R. and Willis, B. L. (1999).** Seasonal and local spatial patterns in the upper thermal limits of corals on the inshore Central Great Barrier Reef. *Coral Reefs*, **18**: 219–228.
- Blakey, T.; Melesse, A. and Hall, M. O. (2015).** “Supervised Classification of Benthic Reflectance in Shallow Subtropical Waters Using a Generalized Pixel-based Classifier across a Time Series.” *Remote Sensing*, **7(5)**: 5098-5116.
- Brown, C. J.; Todd, B. J.; Vladimir, E. K. and Pickrill, R. A. (2011).** “Image-based Classification of Multibeam Sonar Backscatter Data for Objective Surficial Sediment Mapping of Georges Bank, Canada.” *Continental Shelf Research*, **31(2)**: S110-119.
- Browne, N.; Braoun, C.; McIlwain, J.; Nagarajan, R. and Zinke, J. (2019).** Borneo coral reefs subject to high sediment loads show evidence of resilience to various environmental stressors. *Peer J*, **7**: e7382.
- Browne, N.K.; Smithers, S.G. and Perry, C.T. (2013).** Carbonate and terrigenous sediment budgets for two inshore turbid reefs on the central Great Barrier Reef. *Mar. Geol.*, **346**: 101–123.
- Buddemeier, R. W. and Fautin, D. G. (1993).** Coral bleaching as an adaptive mechanism. *Bioscience*, **43**, 320–326. Citizenship and Indigenous Rights. *International Review of Social Research*, **3(1)**: 49-66.
- Coggan, R. and Diesing, M. (2011).** “The Seabed Habitats of the Central English Channel: A Generation on from Holme and Cabioch, How Do Their Interpretations Match-up to Modern Mapping Techniques?” *Continental Shelf Research*, **31(2)**: S132- 150.
- Coles, S. L. and Jokiel, P. L. (1978).** Synergistic effects of temperature, salinity and light on the hermatypic coral *Montipora verrucosa*. *Marine Biology*, **49**: 187–195.

- Coles, S. L., Jokiel, P. L. and Lewis, C. R. (1976).** Thermal tolerance in tropical versus subtropical Pacific coral reefs. *Pacific Science*, **30**: 159–166.
- Darweesh, K. F.; Hallel, A. M.; Saber, S. A.; El Nagggar, H. A. and El Kafrawy, S. B. (2021).** Impact of tourism and fishing on the coral reef health along the west coast of the Gulf of Aqaba, Red Sea, Egypt. *Egyptian Journal of Aquatic Biology and Fisheries*, **25**(4): 785–805.
- Dixon, D.J. (2011).** Tidal Effects on Coral Reefs. In: Hopley D. (eds) *Encyclopedia of Modern Coral Reefs*. Encyclopedia of Earth Sciences Series. Springer, Dordrecht.
- Dixson, D.L.; Abrego D. and Hay M.E. (2014).** Chemically Mediated Behavior of Recruiting Corals and Fishes: A Tipping Point That May Limit Reef Recovery. *Science* **345**: 892–897.
- English, S.; Wilkinson C. and Baker, V. (1997).** *Survey Manual for Tropical Marine Resources*. Townsville, Australia, Australian Institute of Marine Science, Townsville Australia:pp.378.
- Fabricius, K.E. (2005).** Effects of terrestrial runoff on the ecology of corals and coral reefs: Review and synthesis. *Mar. Pollut. Bull.*, **50**: 125–146.
- Flores, F.; Hoogenboom, M.O.; Smith, L.D.; Cooper, T.F.; Abrego, D. and Negri, A.P. (2012).** Chronic exposure of corals to fine sediments: Lethal and sub-lethal impacts. *PLoS ONE*, **7**: e37795.
- Gilmour, J. (1999).** Experimental investigation into the effects of suspended sediment on fertilisation, larval survival and settlement in a scleractinian coral. *Mar. Biol.*, **135**: 451–462.
- Goodkin, N.; Switzer, A.; McCorry, D.; DeVantier, L.; True, J.; Huguen, K.; Angeline, N. and Yang, T. (2011).** Coral communities of Hong Kong: Long-lived corals in a marginal reef environment. *Mar. Ecol. Prog. Ser.*, **426**: 185–196.
- Green, E. P.; Mumby, P. J.; Edwards, A. J. and Clark, C. D. (2000).** Remote sensing handbook for tropical coastal management. *Coastal Management Sourcebooks 3*, UNESCO, Paris.
- Haas, A.F.; Fairoz, M.F.M.; Kelly, L.W.; Nelson, C.E.; Dinsdale, E.A. et al. (2016).** Global microbialization of coral reefs. *Nat. Micro. Biol.*, **1**: 16042.
- Häder, D.P.; Lebert, M. and Helbling, E.W., (2001).** Effects of solar radiation on the Patagonian macroalga *Enteromorpha linza* (L.) J. Agardh—Chlorophyceae. *Journal of Photochemistry and Photobiology B: Biology*, **62**(1-2): 43–54.
- Hamel, M. A. and Andrefouet, S. (2010).** “Using very High Resolution Remote Sensing for the Management of Coral Reef Fisheries: Review and Perspectives.” *Mar. Pollut. Bull.*, **60** (9): 1397–1405.
- Hedely, J. D.; Harborne, A. R. and Mumby, P. J. (2005).** Simple and robust removal of sun glint for mapping shallow-water benthos. *International Journal of Remote Sensing*. **26** (1, part 2): 2107–2112.
- Hedley, J.D.; Roelfsema, C.; Brando, V.; Giardino, C.; Kutser, T.; Phinn, S.; Mumby P.J.; Barrilero, O.; Laporte, J. and Benjamin Koetz (2018):** Coral reef applications of Sentinel 1-2: Coverage, characteristics, bathymetry and benthic mapping with comparison to Landsat 8. *Remote Sensing of Environment* **216**: 598–614.

- Hemdan, G. (1980).** Personality of Egypt, Study of the Intelligence of the place. **4**: s: dār al-ālam al-‘araby. Cairo.
- Herold, M.; Goldstein, N.C. and Clarke, K.C. (2003).** The spatiotemporal form of urban growth: measurement, analysis and modeling. *Remote sensing of Environment*, **86**(3): 286-302.
- Hoegh-Guldberg, O. (1999).** Climate change, coral bleaching and the future of the world’s coral reefs. *Australian Journal of Marine and Freshwater Research*, **50**: 839–866.
- Hoey, A.S. and Bellwood, D.R. (2011).** Suppression of herbivory by macroalgal density: critical feedback on coral reefs? *Ecol. Lett.* **14**: 267–273.
- Hopley, D. (1986).** Corals and reefs as indicators of paleo-sea levels. In van de Plassche, O. (ed.), *Sea-Level Research: A Manual for the Collection and Evaluation of Data*. Ed: Orson van de Plassche. Norwich: Geo Books, pp. 195–228.
- Karkabi, N., (2013).** Lifestyle Migration in South Sinai, Egypt: Nationalization, Privileged Citizenship, and Indigenous Rights. *International Review of Social Research*, **3**(1): 49-66.
- Khaled, M.; Muller-Karger, F.; Obuid-Allah, A.; Ahmed, M. and El-Kafrawy, S. (2019).** Using landsat data to assess the status of coral reefs cover along the Red Sea Coast, Egypt. *Int. J. Ecotoxicol. Ecobiol.* **4**: 17-31.
- Larcombe, P.; Costen, A. and Woolfe, K.J. (2001).** The hydrodynamic and sedimentary setting of nearshore coral reefs, Central Great Barrier Reef shelf, Australia: Paluma Shoals, a case study. *Sedimentology*, **48**: 811–835.
- Lavie, S. (1991).** The Bedouin, the Beatniks, and the Redemptive Fool. *Quar. Rev. of Film & Video*, **13** (1-3): 23-44.
- Lillesand, T. M.; Kiefer, R. W. and Chipman, J. W. (2004).** *Remote Sensing and Image Interpretation*. New York: John Wiley and Sons, Inc.
- Lucas, M. Q. and Goodman, J. (2014).** “Linking Coral Reef Remote Sensing and Field Ecology: It’s a Matter of Scale.” *Journal of Marine Science and Engineering*, **3**(1): 1- 20.
- Macleod, R.D. and Congalton, R.G. (1998).** A quantitative comparison of change-detection algorithms for monitoring eelgrass from remotely sensed data. *Photogrammetric engineering and remote sensing*. **64**(3): 207-216.
- McClanahan, T. R., Maina, J., Moothien-Pillay, R., and Baker, A. C. (2005).** Effects of geography, taxa, water flow, and temperature variation on coral bleaching intensity in Mauritius. *Marine Ecology Progress Series*, **298**: 131–142.
- Mishra, D. R.; Narumalani, S.; Rundquist, D. and Lawson, M. (2005).** Characterizing the vertical diffuse attenuation coefficient for downwelling irradiance in coastal waters: Implications for water penetration by high resolution satellite data. *Photogrammetry and Remote Sensing*. **60**: 48–64.
- MOHPUC (Ministry of Housing, Public Utilities, and Urban Communities), (1997).** Al- (General Plan of the City of Dahab 2017 – Part One: Planning Studies). General Institute of Urban Planning. Urban Planning Center of Suez Canal Region.
- Mona, M.H.; El-Naggar, H.A.; El-Gayar, E.E.; Masood, M.F. and Mohamed, E.S.N. (2019).** Effect of human activities on biodiversity in Nabq protected area, south Sinai, Egypt. *The Egyptian Journal of Aquatic Research*, **45**(1): 33-43.

- Morgan, K.M.; Moynihan, M.A.; Sanwlani, N. and Switzer, A.D. (2020).** Light Limitation and Depth-Variable Sedimentation Drives Vertical Reef Compression on Turbid Coral Reefs. *Front. Mar. Sci.*, **7**: 571256.
- Morgan, K.M.; Perry, C.T.; Smithers, S.G.; Johnson, J.A. and Daniell, J.J. (2016).** Evidence of extensive reef development and high coral cover in nearshore environments: Implications for understanding coral adaptation in turbid settings. *Sci. Rep.*, **6**: 29616.
- MOUNCHU (Ministry of Urbanization, New Communities, Housing, and Utilities), (1993).** Strategic Planning for Developing Sinai. Executive Authority for Urbanizing Sinai.
- Muller, D. and Zeller, M. (2002).** Land use dynamics in the central highlands of Vietnam: a spatial model combining village survey data with satellite imagery interpretation. *Agricultural Economics*. **27**(3): 333-354.
- Nakamura, T. and vanWoesik, R. (2001).** Water-flow rates and passive diffusion partially explain differential survival of corals during the 1998 bleaching event. *Marine Ecology Progress Series*, **210**: 301–304.
- Nakamura, T., Yamasaki, H. and vanWoesik, R. (2003).** Water flow facilitates recovery from bleaching in the coral *Stylophora pistillata*. *Marine Ecology Progress Series*, **256**: 287–291.
- Naumann, M.S.; Bednarz, V.N.; Ferse, S.C.A.; Niggel, W. and Wild, C. (2015).** Monitoring of coastal coral reefs near Dahab (Gulf of Aqaba, Red Sea) indicates local eutrophication as potential cause for change in benthic communities. *Environ. Monit. Assess.* **187**: 44.
- Palmer, S.E.; Perry, C.T.; Smithers, S.G. and Gulliver, P. (2010).** Internal structure and accretionary history of a nearshore, turbid-zone coral reef: Paluma Shoals, central Great Barrier Reef, Australia. *Mar. Geol.*, **276**: 14–29.
- Pizarro, V.; Rodríguez, S.C.; López-Victoria, M.; Zapata, F.A.; Zea, S.; Galindo-Martínez, C.T.; Iglesias-Prieto, R.; Pollock, J. and Medina, M. (2017).** Unraveling the structure and composition of Varadero Reef, an improbable and imperiled coral reef in the Colombian Caribbean., *Peer J*, **5**: e4119. doi:10.7717/peerj.4119.
- Rasher, D.B. and Hay, M.E. (2010).** Chemically rich seaweeds poison corals when not controlled by herbivores. *PNAS* **107**: 9683–9688.
- Reverter, M.; Jackson M.; Daraghme N.; Von Mach C. and Milton, N. (2020).** 11-yr of coral community dynamics in reefs around Dahab (Gulf of Aqaba, Red Sea): the collapse of urchins and rise of macroalgae and cyanobacterial mats. *Coral Reefs* **39**: 1605–1618.
- Richard, K. A. and Ervin, B. H. (2008).** Investigating the effects of higher spatial resolution on benthic classification accuracy at Midway atoll. M.Sc. Thesis, Naval Postgraduate school. Lieutenant, United States Navy.
- Rinkevich, B. (2005).** What do we know about Eilat (Red Sea) reef degradation? A critical examination of the published literature. *J. Exp. Mar. Biol. Ecol.*, **327**: 183–200.
- Robinson, I.S. (2004).** Measuring the Oceans from Space. Chichester, UK: Praxis Publishing LTD.
- Rogers, C.S. (1990).** Responses of coral reefs and reef organisms to sedimentation. *Mar. Ecol. Prog. Ser.*, **62**: 185–202.

- Sanders, D. and Baron-Szabo, R.C. (2005).** Scleractinian assemblages under sediment input: Their characteristics and relation to the nutrient input concept. *Palaeogeogr. Palaeoclimatol. Palaeoecol.*, **216**: 139–181.
- Sarnowski, A.V. (2010).** The role of indigenous people in national development processes: Participation and marginalisation of indigenous Bedouin in South Sinai tourism development, Mainz, PhD Dissertation, Department of Chemistry and Pharmacy, Geosciences, Johannes Gutenberg University Mainz.
- Scoffin, T.P. (1977).** Sea Level feature on reefs in the Northern Province of the Great Barrier Reef. In *Proceedings of the Third International Coral Reef Symposium*. University of Miami, Florida. *Geology*, **2**: 319–324.
- Sheppard, C.R.C. (2003).** Predicted recurrences of mass coral mortality in the Indian Ocean. *Nature*, **425**: 294–297.
- Sherif, E.; Sherif, I. and Abou El-Ela, M. (2016).** political decisions, social change and planning the case of dahab, egypt. In *Proceedings of the 16<sup>th</sup> International City Planning and Urban Design conference (CPUD)*, İstanbul, Turkey. 382-391.
- Singh, A. (1989).** Review article digital change detection techniques using remotely-sensed data. *International journal of remote sensing*. **10**(6): 989-1003.
- Sully, S. and van Woelk, R. (2020).** Turbid reefs moderate coral bleaching under climate-related temperature stress. *Glob. Chang. Biol.* **26**: 1367–1373.
- Swift, S.A. and Bower, A.S. (2003).** “Formation and Circulation of Dense Water in the Persian /Arabian Gulf.” *Journal of Geophysical Research: Oceans*, 108 (C1). The Intelligence of the place). Vol. 4. Publisher: ‘alam al-kutub. Printing press: dār al-‘alam al-‘araby. Cairo.
- Tilot, V.; Leujak, W.; Ormond, R.F.G.; Ashworth, J.A. and Mabrouk, A. (2008).** Monitoring of South Sinai coral reefs: influence of natural and anthropogenic factors. *Aquat. Cons.* **18**: 1109–1126.
- Valentine, P.C.; Todd, B.J. and Kostylev, V.E. (2005).** “Classification of Marine Sublittoral Habitats with Application to the Northeastern North America Region.” *American Fisheries Society Symposium*, Bethesda, Maryland, **41**: 183-200.
- Van Duin, E.H.S.; Blom, G.; Los, F.J.; Maffione, R.; Zimmerman, R.; Cerco, C.F.; Dortch, M. and Best, E.P.H. (2001).** Modeling underwater light climate in relation to sedimentation, resuspension, water quality and autotrophic growth. *Hydrobiologia*, **444**: 25–42.
- Weber, M.; De Beer, D.; Lott, C.; Polerecky, L.; Kohls, K.; Abed, R.M.M.; Ferdelman, T.G. and Fabricius, K.E. (2012).** Mechanisms of damage to corals exposed to sedimentation. *Proc. Natl. Acad. Sci. USA*, **109**: E1558–E1567.
- White, J. and Fitzpatrick, F. (2007).** “How Do I Collect My Sata?” *MESH Guide to Habitat Mapping*. MESH Project, JNCC, Peterborough.
- Williams, S.; Pizarro, O.; Jakuba, M. and Barrett, N. (2010).** AUV Benthic Habitat Mapping in South Eastern Tasmania. In “*Field and Service Robotics*” Howard, A., Iagnemma, K. & Kelly, A. (Eds.) Springer Verlag, Germany, **62**: 275-284.
- Wilson, B.; Blake, S.; Ryan, D. and Hacker, J. (2011).** Reconnaissance of species-rich coral reefs in a muddy, macro-tidal, enclosed embayment, Talbot Bay, Kimberley, Western Australia. *J. R. Soc. West. Aust.*, **94**: 251–265.

- Wilson, E.H.; Hurd, J.D.; Civco, D.L.; Prisløe, M.P. and Arnold, C. (2003).** Development of a geospatial model to quantify, describe and map urban growth. Remote sensing of environment. **86(3)**: 275-285.
- Xiao, J.; Shen, Y. and Ge, J. (2006).** Evaluating urban expansion and land use change in Shijiazhuang, China, by using GIS and remote sensing. Landscape and urban planning. **75(12)**: 69-80.
- Zainal, A.; Dalby, D. and Robinson, I. (1993).** Monitoring marine ecological changes on the east coast of Bahrain with Landsat TM. Photogrammetric Engineering and Remote Sensing;(United States). **59(3)**.
- Zakai, D. and Chadwick-Furman, N.E. (2002).** Impacts of intensive recreational diving on reef corals at Eilat, northern Red Sea. Biol. Cons. **105**: 179–187.
- Zakaria, H.Y.; El-Kafrawy, S.B. and El-Naggar, H.A. (2019).** Remote Sensing Technique for Assessment of Zooplankton Community in Lake Mariout, Egypt. Egypt. J. Aquat. Biol. Fish., **23(3)**: 599-609.
- Zaneveld, J.R.; Burkepile, D.E.; Shantz, A.A.; Pritchard, C.E.; McMinds, R.; Payet, J.P.; Welsh, R.; Correa, A.M.S.; Lemoine, N.P.; Rosales, S.; Fuchs, C.; Maynard, J.A. and Thurber, R.V. (2016).** Overfishing and nutrient pollution interact with temperature to disrupt coral reefs down to microbial scales. Nat. Comm. **7**: 11833.
- Zewlifer, A., O'Leary, M., Morgan, K. and Browne, N.K., (2021).** Turbid Coral Reefs: Past, Present and Future—A Review. Diversity, [online] **13(6)**: 251.

### الملخص العربي

ثلاثة عقود من مراقبة التغير في البيانات القاعية باستخدام الاستشعار من البعد على الساحل المصري لخليج العقبة، البحر الأحمر

كريم فاروق درويش<sup>١</sup>، أحمد متولي هلال<sup>١</sup>، سامي عبد اللطيف صابر<sup>١</sup>، سامح بكر الكفراوي<sup>٢</sup>، علي عبد الحميد عبد السلام<sup>٢</sup> و حسين عبد المجيد النجار<sup>١</sup>

<sup>١</sup> قسم علم الحيوان، كلية العلوم، جامعة الأزهر، القاهرة، مصر  
<sup>٢</sup> قسم علوم البحار، الهيئة القومية للاستشعار من البعد وعلوم الفضاء، مصر

شهد ساحل البحر الأحمر خلال الأربعة عقود الأخيرة تغيرا ونموا كبيران. في خلال فترة قصيرة من الزمن تحول ساحل خليج العقبة الي سلسلة من المنشآت السياحية. نظرا للندرة الشديدة في المعلومات الخاصة بتلك المنطقة، في هذه الدراسة تم وضع تحليل زمني شامل لتحديد التغيرات باستخدام تقنيات الاستشعار من البعد لهذه المنطقة. تم استخدام البيانات المستخرجة من القمر الصناعي (Landsat) في خلال الفترة من ١٩٨٥ الى ٢٠٢٠ في منطقة خليج العقبة فيما يخص تحديد التغيرات في المنشآت و خط الشاطئ و الفترة من ١٩٩٥ الي ٢٠٢٠ في تحديد التغيرات المتعلقة بتوزيع البيانات القاعية كما تم استخدام الصور من Landsat 5-TM, Landsat 7-ETM+ Landsat 8-OLI في وضع سلسلة من مشاهدات الأقمار الصناعية. أظهرت النتائج تغيرا ملحوظا في الثلاثة عقود الأخيرة في قطاع دهب و قطاع نويبع بواقع ٤.١ و ٢.٨ كم مربع على الترتيب، بينما كان التغير في خط الشاطئ في نطاق أقل من سابقه بشكل منفي و من دون احداث الضرر للموارد الطبيعية. تبين ان التغيرات في كلا من المنشآت وخط الشاطئ تتناسب طرديا مع الزيادة في النشاط السياحي. فيما يخص التحليل الزمني العامل الفيزيائية البيئية كان هناك اختلافات واضحة خصوصا في الفترة الزمنية من ١٩٩٣ الى ٢٠٠٣ في كلا من ارتفاع سطح الماء و العكارة بينما كانت درجة حرارة سطح ماء البحر اتجاه ثابت مستقر. توزيع البيانات القاعية الزمني اظهر ثبات في معظم البيانات في قطاع دهب ماعدا زيادة ملحوظة في بيئة الشعاب المرجانية في الفترة من ١٩٩٥ الى ٢٠٠٥، كما كان هناك تقلب في توزيع الشعاب المرجانية في قطاع نويبع.